

Wetlands and mosquitoes: a review.

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Abstract

This review brings together information on mosquitoes, the diseases they transmit and the wetlands that provide habitats for the immature stages (eggs and larvae). Wetland values are mentioned, though the main literature on this does not generally overlap the mosquito issue. Mosquito management is overviewed to include: the use of larvicides, source reduction in intertidal wetlands and management in freshwater systems. There is not a great deal of information on mosquitoes and freshwater systems, except for constructed wetlands and they are considered separately. We then consider restoration mainly in the context of wetlands that have been the subject of habitat modification for mosquito control. Land use and climate change, as they affect mosquitoes and the diseases they transmit, are also reviewed, as this will affect wetlands via management activities. Finally the review addresses the critical issue of balancing health, both human and environmental, in an adaptive framework. It concludes that there is a need to ensure that both mosquito and wetland management communicate and integrate to sustain wetland and human health.

Keywords: freshwater wetland; intertidal wetland; mosquito; mosquito-borne disease;
mosquito control; wetland

Introduction

Wetlands are indisputably important ecosystems, as evidenced by a wealth of literature spanning several decades. Many conferences, books and papers have been devoted to the subject of wetlands: their importance, threats to their integrity and management issues. There is also an extensive literature on mosquito-borne disease, mosquitoes and their management. The immature stages of mosquitoes need water and so there is a coincidence of wetlands and mosquito-borne disease vectors at both the global and local scales. To illustrate this at the global scale Figure 1 shows the world distribution of wetlands of various kinds and the distribution of some of the major mosquito-borne diseases and their vectors: malaria, yellow fever, dengue, filariasis and Japanese encephalitis. At the scale presented, details are not shown for the salt-affected or inland wetlands nor are the large areas of permafrost in the northern hemisphere shown, though global warming could reduce the extent of permafrost and replace it with other wetlands.

Insert Fig. 1

This review focuses on the issue of wetlands, mosquitoes and mosquito-borne disease. It will first provide an overview of mosquito ecology, focussing on the habitats of the immature stages. It will then very briefly consider the wetlands of the world and their values, as these values may be impacted by mosquito management. The major mosquito-borne diseases will be reviewed including references to the wetland habitats that are

important to them. The review will cover mosquito management and its impact on a range of wetlands, impacts of change (especially of climate change), and adaptive approaches to balancing the environment and human health issues. It will also highlight gaps in knowledge that need to be filled in order to understand the issues and optimise wetland and disease-vector mosquito management.

Approach

The topic of wetlands and mosquitoes is a large one. In order to obtain an overview of the field, as well as key references, we carried out a broad preliminary analysis of the literature. The database was compiled in Endnote v 9 and comprised searches on title, keywords and abstract content, to provide a broad view. The wide search was used to inform specific aspect of the review but frequency counts were restricted to title content, for focussed information. Databases included Amed and Healthstar (both via OVID), Pubmed, Web of Science and the results were combined to include the authors' personal Endnote records. Preliminary sorting removed duplicates and items that, although important to the area were not the focus of the current review. These included *inter alia* genetic analyses of mosquitoes (though not if also related to management), pathogens, vaccine development and clinical case studies. A database of over 3000 references resulted and was sorted by categories, such as 'mosquito' AND 'wetland', 'disease' AND 'mosquito'. We tried to include as many review papers as possible since they represent a synthesis that helps cover at least indirectly the breadth of the topic area. In the following sections we first outline results showing how the literature has changed since 1983 and how focus has shifted within sub-categories. We chose 1983 as it is convenient for

searching electronic databases and, in any case, important earlier papers are referenced in the papers that we have cited here.

Mosquitoes and wetlands

The post 1983 literature shows a recent large increase in interest in mosquitoes and wetlands and highlights the increasing recognition that both aspects are important. Around 10% of the references included as terms both mosquito and wetlands of various types (e.g., salt marsh (in its various forms), mangrove, swamp etc). Of these 21% were published recently, between 2004 and 2007. The following section summarises mosquito ecology with emphasis on the role of wetlands as mosquito habitats.

Mosquito ecology overview

Mosquitoes are arthropods (in the Phylum Arthropoda). The virus diseases they can transmit are often referred to as arboviruses (arthropod-borne viruses). Mosquitoes belong to the Class Insecta, the Order Diptera (flies) and, within the order, all mosquitoes belong to the Family group called Culicidae. Within the family there are Sub families and then Genera. Mosquito genera important to human disease transmission include *Aedes*, *Culex* and *Anopheles*. Each has distinct habitat requirements, and there are differences between species. The general pattern of mosquito development is from egg to hatch through four larval instars, to pupation and then emergence as adults. The timeframe for a complete cycle may be as short as five days in tropical and subtropical environments and there are differences between species in their development times.

All mosquitoes have an intimate relationship with wetlands. Water is an essential requirement for the larval stages. There are many reference works on mosquito ecology. An early reference work is that of Lounibos et al. (1985) covering community and population dynamics, ecology, epidemiology and also the role of genetics in the life strategies of the insect. The major reference text remains that of Service (1993) who reviewed the literature on the topic of the ecology, sampling and modelling of mosquito populations. There is also a literature that is used by mosquito control agencies, on a national or regional basis, such as the work for south east Australia by Russell (1993) that provides a concise overview of the ecology of and key to the common mosquitoes and their typical habitat characteristics.

The eggs are laid in water in rafts of multiple eggs (*Culex* spp), on water singly (*Anopheles* spp) or singly on damp substrate that will later be flooded (*Aedes* spp). In water, the eggs hatch into larvae that go through four instar stages. During this stage they feed on small organisms or decaying material. They breathe air through a siphon that is at or protrudes through the water surface, or, in some species, they attach to plant stems and obtain oxygen directly from the plant tissue. After the fourth instar they pupate and then emerge as adult flying insects. In some species the newly emerged adult female may be able to lay fertile eggs but generally a blood meal is required for protein to produce eggs. For disease vectors the pathogen is picked up during the blood meal and, if it replicates within the mosquito, it may be transmitted later to a victim during another blood meal. In some cases diseases may be passed on directly from the adult female via the egg to the larva and hence to the emerging adult. This is known as vertical transmission. From a wetland perspective it is important to identify the habitats of the immature stages (eggs

and larvae) as these habitats are usually wetlands and the focus of larval, and hence wetland, management.

Oviposition sites (egg laying)

Although the egg stage determines generally the area that larvae will start their cycle, information is not detailed for all vectors of disease. Knowing where oviposition sites are is an aid to identifying larval habitats as larval survey may miss some of these, such as ephemeral sites when they are dry. Mosquitoes that have received much attention include the aedine nuisance species or vectors of viruses, such as *Aedes taeniorhynchus* (Wiedemann) in Florida and *Aedes vigilax* (Skuse) in Australia. That is, in part, because these species lay eggs on the ground surface and hence can be sampled directly (as compared to setting oviposition traps). Eggshells of *Aedes* spp are good indicators of oviposition and can be used to sample at times when larvae are not abundant or are absent, as eggshells are relatively stable both spatially and temporally (Ritchie 1994). Eggshell studies have been carried out for open vegetation such as salt marsh, mainly in Australia (Dale et al. 1986; Kay and Jorgensen 1986; Ritchie 1994; Ritchie and Jennings 1994; Gislason and Russell 1997; Turner and Streever 1997; Dale et al. 2008) but there has been relatively little research on the more complex mangrove forest systems (Ritchie and Johnson 1991b; Ritchie and Johnson 1991a; Ritchie and Addison 1992).

Mosquito larval habitats

The element common to all mosquito species is the need for water for the larval and pupal stages. Knowing the association between wetlands and mosquitoes is important for mosquito-borne disease as it helps to focus management. Rodriguez et al. (1996) found

associations between landscape features including elevation and vegetation (mangroves) and malaria vectors in Mexico. Mercer et al. (2005) found significant relationships between immature mosquitoes and habitat characteristics such as water quality, in wetlands in Iowa (USA). They also explored the literature to further identify microhabitats associated with mosquitoes and used these to estimate the risk of disease transmission. Most of the published research on larval habitats of *Aedes taeniorhynchus* in mangrove forests has been by Ritchie and colleagues, for example Ritchie and Addison (1992). Although not focussing on larvae, there have been compilations of mosquito species recorded, for instance, in Indian mangrove systems (Rajavel et al. 2005a; Rajavel et al. 2005b; Rajavel and Natarajan 2006).

At a detailed level, Sogoba et al. (2007) in Mali, for malaria vectors, found that dry season habitats provided a refuge for mosquito populations and suggested that focusing larval control on those would reduce the wet season populations. Miller et al. (2007) showed that *Anopheles gambiae* s.l. larvae can survive in wet mud as well as in water and that this explained their survival in an uncertain environment (with implications for management).

Detailed local studies of larval habitats highlight the need to efficiently identify them so as to focus management (Balling and Resh 1983). Field survey is important to confirm the larval habitat status of a site but Remote Sensing (RS) and Geographic Information Systems (GIS) can speed up the process and survey large areas cost effectively. Dale et al. (1998) reviewed some examples for disease vector mosquitoes in Australia. Other research includes Dale and Morris (1996) who used aerial photographs to identify ephemeral wetlands in urban subtropical Australia; Mushinzimana et al. (2006) who used

1 m spatial resolution satellite imagery for African highlands in Kenya and Bian et al. (2006) who used GIS to estimate the spatial distribution of anopheline larval habitats in the highlands of West Kenya, taking into account terrain, landuse and surface water. Several studies have mapped the larval habitats of vectors of West Nile virus (WNV) in USA. For example, Diuk-Wasser et al. (2006) developed a spatial model showing that various mosquito vectors had clear habitat preferences and Zou et al. (2006) were able to distinguish larval habitats in Wyoming (USA) using Landsat imagery.

Information gaps

Detailed information on the immature stage habitats is relatively sparse in the refereed literature. Most of the local information is held, if at all, by vector control agencies. Management needs to have information at a scale relevant to the organism and that is at a fine level of detail. This information gap was noted by Vezzani et al. (2006) who saw a need to identify larval habitats of filariasis vector mosquitoes so as to focus management on relevant areas. In particular, there is lack of research on the immature stages of mosquitoes in forested wetlands, both freshwater and intertidal. Such research is inhibited by survey difficulty but there is potential to use novel RS technology (e.g., Dale et al. 2005b), though resources may limit this.

Wetlands of the world and their values

That wetlands are important and valuable is widely acknowledged. Schuyt and Brander (2004) have documented wetland types and their economic values and this is

summarised at a global scale in Table 1.

Insert Table 1

According to Schuyt and Brander, North America and Latin America contain significant proportions of the world's wetlands (36% and 19% respectively) as well as having serious mosquito-borne diseases.

That wetlands are threatened by change both natural and anthropogenic has been recognised in books and compilations such as that of Mitsch (1998) who noted the importance of the world's wetlands and the processes that threaten them. A recent article in Science articulated the concern that mangrove losses threaten the ecosystem services they provide and thus the resource for future generations (Duke et al. 2007). A significant global advance in documenting the world's wetlands of international significance was made by the Ramsar convention in 1971. There was a recent overview of this and comprehensive inventory of listed wetlands as of March 2008, when 160,158,832 hectares of wetlands in 1722 sites were protected by the convention (Secretariat of the Ramsar Convention 2008). The Ramsar aim is to develop and extend a global list of wetlands that are ecologically significant for maintaining biodiversity, ecological functions and for sustaining systems upon which much life depends.

In the overlap between mosquitoes and wetlands, wetland value is not often mentioned and mosquito value is itself a novel concept. Exceptions include Mercer et al. (2005) who stressed the value of wetlands and the ecological services they provide, measuring the abundance of mosquitoes as well as other insects and recording other

environmental variables such as water quality and vegetation cover in freshwater reconstructed wetlands in Iowa USA. Mosquito value was also addressed by Schafer (2004) who considered the contribution that mosquitoes make to wetland biodiversity in a range of wetland types from forest to meadow to shallow water. Schafer's research is relevant to addressing a frequently asked question: What use are mosquitoes? Mokany (2007) suggested that mosquitoes, being abundant in ephemeral wetlands, can have a significant effect on ecosystem processes and functions and, by implication, wetland value.

A great deal of the wetland literature is in the area of ecology, hydrology and ecohydrology, with the latter as an important area that brings the first two disciplines together. However, relatively few papers also overlap with mosquitoes or their habitats although most wetlands have some potential as mosquito habitat. To manage the mosquito problem the ecohydrology of their environment needs to be understood. References that are useful in this respect for the intertidal systems in general include Elliott et al. (2007). For specific systems, the book by Mazda et al. (2007) brings together a literature that addresses mangrove processes in detail in order help conserve the systems. The review of salt marsh management by Dale and Hulsman (1990) addressed both salt marsh values and mosquito management issues. A specific study demonstrating the application of an hydrological approach is that of (Shaman et al. 2002) who used a dynamic hydrology model, based on meteorological data, topography, soils and vegetation to model wetness and thus to predict mosquitoes in New Jersey USA. Freshwater wetlands are covered to some extent by the constructed wetland literature (reviewed below) but there appears to be little information on freshwater forested

wetlands.

Information gaps

Wetland studies that consider wetland values, from a human perspective, tend to focus on the positive values and do not take much account of the negative ones. This needs to be addressed through research to inform a balanced approach to management. Having made that point, there is also an argument that intrinsic value should include the fauna in its entirety, including mosquitoes, as contributing to value, demonstrated, for example, by Schafer (2004). This again needs more research to investigate the nature of the values.

Mosquito transmitted diseases

There has been a growing literature on mosquitoes and disease, especially for WNV, after its introduction into USA in 1999, affecting both humans and animals. Of over 1000 references to WNV in our database, 96% were dated 1999 or later. The spread of WNV prompted the review of host-vector disease models by Wonham et al. (2006). That 78% of all the references relate to WNV and diseases such as Ross River virus also highlights the preponderance of research into diseases of the developed areas of the world that have much less serious consequences for human health than, for example, malaria or filariasis. There are also general reviews for arboviruses in particular regions, such as the review of a wide range of arboviruses in western Europe by Lundstrom (1999). Vector competence (the ability to transmit disease) may be used to prioritise mosquito and wetland management and so is briefly introduced in the next section.

Vector competence

Mosquitoes may be a nuisance, but where they are also capable of transmitting disease to human (or animals), the case for direct or indirect management becomes important. Although there may be a correlation between the presence of specific mosquitoes and mosquito-borne disease, vector competence research is needed to confirm the relationship, usually by means of laboratory experiments. There is a large amount of research on this area, some of which reports correlations between specific diseases such as WNV and mosquito species (e.g., Dennett et al. 2007). Most vector competence research focuses on a specific mosquito or on a specific disease. A few recent examples are given here. For a mosquito species *Aedes aegypti* (L.) has received much attention: as a vector of filariasis (Tiawsirisup and Nithiuthai 2006); of yellow fever (Johnson et al. 2002; Jupp and Kemp 2002) and of dengue (Knox et al. 2003). For a specific disease, WNV, the vector competence of several species has been explored, for example of *Culex* spp (Weng et al. 2000; Vaidyanathan and Scott 2007; Balenghien et al. 2007; Reisen et al. 2006), of *Culiseta* and *Culex* spp (Reisen et al. 2006) and of Californian species including *Culex* spp, *Ochlerotatus* spp and *Aedes vexans* (Meigen) (Goddard 2003; Chen et al. 2000). The next section provides an overview of the major diseases shown in Fig. 1 (and some others) and the wetland environments that support their vectors.

Malaria

Malaria in humans is a serious disease, endemic to the tropics, whose pathogen (*Plasmodium* spp) is transmitted by mosquitoes of the genus *Anopheles*. The mosquitoes

occupy a wide range of habitats from freshwater to saline and these may be subdivided according to elevation, landform or land use. The literature includes area-specific research on the *Anopheles* spp habitats. Some, that also focus on disease, are shown in Table 2 and others that focus on the mosquitoes include Stoops et al. (2007) who found general associations with higher elevation and lower tree canopy coverage, higher water temperatures, and shallow water as well as more species specific associations in Indonesia. Muturi et al. (2006) found relationships between land use and mosquito abundance (*Anopheles* and *Culex* spp) in Kenya. Climate, especially temperature, is important as the pathogen does not thrive in low ($<16^{\circ}\text{C}$) or high ($>32^{\circ}\text{C}$) temperatures. This has implications for climate change and its potential impacts on malaria distribution.

Malaria in Sub-Sahara Africa has received a great deal of attention, as it is one of the most significant diseases there. It is also significant in other places and a review of research conducted in six areas in Indonesia that examined the relationships between malaria and environment, including socio-economic variables, found spatial and temporal heterogeneity in most variables, indicating the need for locally specific information for management (Dale et al. 2005a).

Some of the genetic research also has relevance to the distribution of the disease and hence control of wetland vectors. For example a recent paper by Yawson et al. (2007) concluded that ecological barriers are more important to gene flow in *Anopheles gambiae* than geographical distance. Different genotypes may have different vector competencies resulting in different disease incidences and hence priorities for mosquito or wetland management.

Yellow fever

Yellow fever is not well covered in the research literature, at least for material that includes wetlands. Much research appears to be in the area of genetics of its major vector *Aedes aegypti* and hence is not covered here. As well, the vectors are mainly container breeders and natural wetlands are not necessarily the preferred habitat. A useful review is that of Bourgeade and Marchou (2003) who describe the epidemiology of yellow fever in Africa and South America.

Filariasis

Filariasis is vectored by a range of mosquito species both aedine, anopheline and culicine and infects both humans and dogs. It occurs widely (Fig. 1) but is not much reported in terms of the focus of this review. Vezzani et al. (2006) review its history in Argentina. Its vectors in Italy were reviewed by Cancrini et al. (2006).

Japanese encephalitis

Japanese encephalitis (JE) is a virus transmitted by *Culex* spp. and often associated with pigs close to human habitation. It may be included in discussion of WNV, though JE is a more serious disease, often resulting in death or permanent impairment. It is only found in Asia in a rough triangle from Pakistan to Indonesia and to Siberia (Spira 2007) though its vectors are more widespread (Fig. 1). Bourgeade and Marchou (2003) showed three zones of the disease: epidemic, to the north (e.g., Korea, Japan, Nepal, northern India), endemic to the south (Papua-New Guinea, southern India, Indonesia) and a

transition zone between. There are many papers on aspects of the disease itself that also include the term mosquito but few refer to wetlands. A study by Hemmerter et al. (2007) is an example of genetic research relevant to this review. They suggested that the lack of JE in Australia may be related to a different lineage of *Culex annulirostris* (the major vector in south east Asia), and it may not be a very competent vector.

West Nile virus.

Since the arrival of WNV in USA in 1999 it has generated a great deal of interest. Initially it was not regarded as a very serious disease but it is now recognised as causing death, especially in the elderly (Bourgeade and Marchou 2003). Kramer et al. (2007) reviewed the disease comprehensively and included information on its ecology. Petersen and Roehrig (2001) suggested that WNV was a re-emerging global pathogen and that theme has been developed by others, including, for example, Boyer et al. (2002) who viewed WNV as the first pandemic of the 21st century. Others include Gerhardt (2006), Glaser (2004), Hayes et al. (2005) and Hayes and Gubler (2006).

Concern over the spread of WNV was reflected by Hubalek (2000), who reviewed its epidemiology in Europe with the aim of informing what might happen in USA after the initial detection of the virus in 1999. There have been several reviews of aspects of WNV such as its introduction and subsequent spread in USA (Gould and Fikrig 2004), its biology and potential to spread to South America, citing mosquitoes in the genus *Culex* as the major vector (Granwehr et al. 2004). Jourdain et al. (2007) referred to risk in the Camargue in southern France and Kaptoul et al. (2007) noted the first case in Spain. Risk

of WNV has been assessed for the British Isles by Medlock et al. (2005), based on the ecology of the mosquito vectors.

Many of the references relate to birds: their deaths and their potential to spread the disease e.g., Medica et al. (2007), Koenig et al. (2007) and Ladeau et al. (2007). Hubalek (2000) noted that wetlands are implicated in the transmission cycle via the role of birds.

As well as indicating mosquito vectors some papers also provide information on their habitat characteristics, landscape and climate. Pecoraro et al. (2007) found relationships between WNV and landscape and climate (this is also relevant to climate change concerns, to be discussed later).

Ross River and Barmah Forest viruses

Ross River virus is important in the Asia Pacific and especially in Australia. It has been reviewed by several authors including Russell (2002) and Gatton et al. (2004). It is widely distributed, as there are several vectors whose habitats cover a range of wetland type from intertidal to freshwater and from permanent to ephemeral water bodies (Kelly-Hope et al. 2004). Muhar et al. (2000) found that risk of Ross River virus in Brisbane, Australia was significantly related to several wetland habitats including intertidal and freshwater ones associated with floodplains. More recently, Barmah Forest virus has become a public health issue in Australia, with symptoms similar to Ross River virus and the two diseases have some mosquito vectors in common (Doggett et al. 1999; Quinn et al. 2005).

Dengue

Dengue is mainly vectored by peridomestic mosquitoes, including *Aedes aegypti* (and more recently *Aedes albopictus* (Skuse)) so it is not a major focus for wetland issues. Nevertheless *Aedes albopictus* vectors other arboviruses (La Crosse, WNV), is also found in natural wetlands and has the ability to spread rapidly, as noted by Benedict et al. (2007).

Information Gaps

What emerges from a review of mosquitoes and disease is that relatively little emphasis is placed on the wetland habitats in which the vectors spend the critical larval stage of development and on which management often concentrates. The focus tends to be on the species that transmit disease and the relationship between vector and host (including intermediate hosts).

In practice, mosquito-borne disease management includes a range of actions from health education and clinical practice to wetland management with a focus on larval habitats. The next section focuses on larval management in wetlands.

Mosquito management and its impacts on wetlands

While this section focuses on mosquito management, the other side of the coin needs to be kept in mind: wetland management and how this interacts with mosquito control. An Australian perspective was provided by Dale and Morris (1994). In the present review 32% of the 430 ‘mosquito’ AND ‘management’ OR ‘control’ references referred to wetlands and, of these, 17% related to constructed wetlands (discussed below). Reducing

the risk of mosquito problems is most effectively done at the larval stages when the larvae are spatially concentrated. Adulticiding may be carried out but, by the time the adult flying stage is reached, dispersal is often several kilometres and may be up to 50 km for some species. It is important to note that larvae *per se* are not a human disease or nuisance problem. The problem is when they emerge as adults ready to take a blood meal and both larviciding and source reduction aim to prevent adult emergence.

There seems to be a trend away from attempting to annihilate mosquitoes or destroy the environment. Had we looked further back in time we would find that early mosquito agencies in USA often had the term ‘extermination’ in their name. For example, the New Jersey Mosquito Extermination Association was created in 1914 and later changed ‘Extermination’ to ‘Control’, as it now is. There is a large grey literature, including manuals for mosquito management published by various mosquito control organisations but these are not reviewed here. Some were referenced in Dale and Hulsman (1990) in the salt marsh context. They noted an evolution from an approach that may destroy the intertidal wetland as well as its mosquito population to a more benign approach that seeks to minimally alter the environment, based on the premise that relatively subtle changes may affect the mosquito without destroying wetland function.

Mosquito management methods have been developed over a lengthy period, especially in USA. Patterson (2004) reviewed the development of mosquito control in USA in his book “Mosquito Wars”. In it he described the evolution from reliance on chemicals such as DDT to source reduction (habitat modification of wetlands) and to the chemicals that are currently used and that are believed to be less injurious to the environment than those used in earlier decades. Floore (2006) reviewed the history specifically of larval control

in USA from the 1900s, when water bodies were treated with kerosene or an arsenical product, to today's methods of larviciding and source reduction.

In the context of disease control there is also research being conducted on ways to reduce the vector competence of mosquitoes, such discussed in Joardar (2005), or to develop methods based on genetics, an approach reviewed by Tu and Coates (2004) and reviewed and explored by Linser et al. (2007) or to use radiation (e.g., Helinski et al. 2006). If successful and adopted they would minimise the direct impact of mosquito management on wetlands. However, none of these are yet in common use and so are not considered in the following sections.

The following will provide a general overview of current larval management methods and their impact on wetlands, to include: source reduction in intertidal wetlands (salt marsh and mangroves), freshwater and forested wetlands (not mangroves) and constructed wetlands.

Larvicides and their impacts on wetlands

Larvicides have a role to play in disease reduction. For example, Walker and Lynch (2007), in their review of larvicide use in tropical Africa, noted its effectiveness in reducing malaria, especially when used in conjunction with other methods (e.g., bednets, source reduction).

Larvicides are generally highly effective at killing mosquitoes, though resistance may become a problem. It is beyond the scope of this review to review the products used to control/manage larval populations but some general comments can be made. There is a range of chemical and biorational larvicides currently in use. The main chemical

categories include, for example, some organophosphates, surface agents such as monomolecular surface films and oils as well as products that are termed biorationals. Floore (2006) listed products that are generally no longer used in USA and reflecting the growing concern for environmental impacts. Those products included various oils, organophosphate and chlorinated hydrocarbons, though specific products varied by state. For biorationals, Floore (2007) edited a 330 page supplement to the Journal of the American Mosquito Control Association dedicated to this topic. The publication contains reviews on a wide range of biorational control agents, a few of which are referenced here. They include the widely used synthetic bacterial agents such as *Bacillus thuringiensis* var *israelensis* (*B.t.i.*), *Bacillus sphaericus* (Lacey 2007) and the insect growth regulator methoprene (Henrick 2007).

Forays into the field of biological control agents have had variable results. Other insects that are predators of mosquito larvae are reviewed by Quiroz-Martinez and Rodriguez-Castro (2007). Copepods are an example of biological organisms that predate on mosquito larvae (Marten and Reid 2007). Fish may be effective in some circumstances, though the use of *Gambusia* may have adverse effects on native fish (Walton 2007). Recent experiments with a mould (*Lagenidium*) were encouraging, including its safety for non-target organisms (Vyas et al. 2007).

Because of resistance problems there is a wealth of literature on testing new products or variants of existing ones. For example Park et al. (2005) developed a new strain of *B.t.i.* effective against vectors of WNV. In the present review over 10% of papers referred to efficacy testing of mosquito control agents but few of these also considered impacts on non-target organisms, an important aspect for wetland health. Exceptions include Vyas et

al. (2007) noted above, for a mould, and Merritt et al. (2005) for *Bacillus sphaericus*, noting no detrimental effects on, for example, taxa richness and abundance of non-target invertebrates, for six applications of insecticide over a three-year study.

Impacts

Dale and Hulsman (1990) in the context of salt marsh mosquito management noted that research on the impacts of chemicals on non-target organisms generally did not assess long-term or chronic impacts. Papers that assess the impacts of larvicides on non-target organisms may include field components associated with laboratory assessment, but long-term effects and effects on food webs are still not well understood. A study of the sub-lethal effects of selected larvicides on a native fish species was investigated by Hurst et al. (2007) and found no significant impacts for microbial larvicides and insect growth regulators, but did note effects of some organophosphate compounds. All were tested at concentrations greater than those used for mosquito management.

Source reduction and its impact on intertidal wetlands

Source reduction simply means reducing the source of mosquito populations. This does not necessarily mean destroying the wetland, as relatively small changes may adversely impact the mosquito life cycle and prevent adult emergence. In the intertidal environment the general group of source reduction methods currently in use in the USA include tidal recirculation, reported by Resh and Balling (1983) for California, Open Marsh Water Management (OMWM), first developed in New Jersey in the late 1960s (Ferrigno and Jobbins 1968), and developed on east coast USA intertidal wetlands,

usually for those that had been ditched. By the 1980s, the OMWM method was widely used and guidelines for its implementation were developed and accepted by the relevant federal agencies (US Corp of Engineers and the US Fish and Wildlife Service). Other guidelines include those of Bruder (1980), Meredith et al. (1985) and Hruby and Montgomery (1986). Open Marsh Water Management was reviewed by Wolfe (1996) and its much smaller Australian version, runnelling, was reported in Hulsman et al. (1989).

In Florida, a state that was extensively impounded for mosquito control in the 1950s and 60s, the recognition of wetland values led to the Rotational Impoundment Management (RIM) concept whereby tidal circulation is allowed during the winter when mosquitoes less of a problem, and the impoundments are flooded up in spring to reduce the summer habitats of the immature stages (Carlson and Carroll Jr. 1983; Carlson and Vigliano 1985; Carlson 1986). The RIM method works by keeping water levels high and consistent and so the damp-drying-wetting conditions needed by aedine mosquitoes for oviposition, egg conditioning and hatch are not satisfied. Schmalzer (1995) reviewed the history of this aspect of the Florida marshes.

Impacts

Impacts of the various methods of source reduction have been assessed, generally over the short term (less than 3 years). Examples include, for OMWM, early papers reported in the reviews by Dale and Hulsman (1990) and Wolfe (1996) and in the assessment by Meredith et al. (1983). Much of the early work was published in conference or meeting proceedings and is not easily accessed (e.g., Romanowski and

Risch 1986). Research was at that time in response to the growing concerns about coastal wetlands, their values and threats to them or to their wildlife, such as birds (Brush 1986). More recent research/monitoring is in reports such as the comprehensive assessment for Suffolk County (Wertheim Demonstration Project) north east USA (Cashin Associates 2007) and theses (Latchford 1997; Breitfuss 2003), and is generally relatively short term. For a longer term study the 20 year monitoring of runneling in Australia is relevant (Dale 2007). Some innovative multivariate impact assessment techniques have been developed for that research examining effects on process rather than simply on the state of the system (Dale et al. 2002; Dale and Dale 2002).

The impacts of impounding have been discussed in the reviews referred to above (e.g., Dale and Hulsman 1990; Patterson 2004). Florida (USA) features largely in this area of research with impacts assessed for a variety of organisms (e.g., for fish, Harrington Jr. and Harrington 1982). Brockmeyer et al. (1997) in their paper on rehabilitation of impounded estuarine wetlands noted that impounding had had “a profoundly negative impact” on the wetlands (p96). A response has been to develop the RIM method (referred to above) that restores or partly restores tidal circulation to impoundments. Impact assessment of RIM has focussed on the extent to which it restores flora and fauna, but a major objective is also to manage mosquitoes (see Carlson and Vigliano 1985). One difficulty is that, after being impounded for some decades, the marsh elevation may not have evolved in line with natural marshes, and may be relatively lower in some impoundments (Parkinson et al. 2006). They suggested that restoring tidal flooding, unless this is done very gradually, may result in water to depths that inhibit plant colonisation and hence jeopardise restoration success.

Freshwater wetlands

Research on freshwater wetland communities may note mosquitoes, but may not mention disease risk or suggest management. For example, Reese and Batzer (2007), studying the floodplain of the unregulated Altamah river in south east USA, found a diverse assemblage of aquatic fauna and mentioned rapidly developing organisms including mosquitoes associated with the upper reaches. Ephemeral wetlands may be important for mosquito populations. Certain species (such as *Culex annulirostris* in Australia) are known to rapidly colonise ephemeral wetlands. In their work on the wetlands of Iowa USA, Mercer et al. (2005) found that the greatest risk of disease transmission was associated with ephemeral pools, an observation supported by others such as Dale and Morris (1996). Also in an ephemeral pond context, Mokany (2007) considered the impact of mosquito larvae (and tadpoles) on the invertebrate community, although without reference to potential human health impacts of the mosquitoes after emergence from the wetland.

Natural freshwater wetlands include both forested wetlands (swamps) and marshes and have been a neglected area of mosquito-specific research. Forested wetlands include habitats for mosquito vectors of disease. For example Hachiya et al. (2007) noted that the areas sampled for the Eastern Equine Encephalitis vector (*Culiseta melanura* (Coquillett)) were in forested wetlands. Eastern Equine Encephalitis can also infect humans. Schafer and Lunstom (2001) compared mosquito species trapped in four forested wetlands in Sweden. There is a literature on managing forested wetlands, which, although not specifically for mosquito control, may be relevant to mosquito issues. An

example is the review by Sun et al. (2001) on the hydrologic impacts of common forest uses, including drainage. Drainage has, particularly in the past, also been used as a source reduction technique for mosquito management. However the research also showed effects such as raised water tables following harvest in some situations, for example in Cypress and other forested swamps and, incidentally, a habitat of *Culiseta melanura* (Hachiya et al. 2007). These are very relevant to the creation of, or increase in, mosquito larval habitats, although not surprisingly this is not mentioned in specifically forest-focussed research.

One method of managing freshwater mosquitoes, apart from larviciding, is source reduction and this may involve draining or filling the larval habitats as described by Shililu et al. (2007). Their work was on malaria vectors in a semiarid part of Africa (Eritrea) and, because of the severe health risk, local communities were enlisted to monitor and fill or drain ephemeral wetlands, though this would destroy wetland function. In that case the human health risk was a paramount consideration. Gu et al. (2006) showed theoretically that source reduction, by reducing oviposition sites, could lead to a longer mosquito life cycle (mosquitoes have to search for longer to find suitable sites to lay eggs) and that this could have important implications for reducing disease incidence (in that case malaria).

Constructed wetlands are a specific category of freshwater wetlands and these have generated a great deal of interest as discussed below.

Constructed wetlands

Constructed wetlands are increasingly being used for water treatment (stormwater, secondary treated effluent) and also contribute wetland values to the environment.

Schafer et al. (2004) assessed biodiversity in natural and constructed wetlands in temperate Sweden. They found mosquito abundance and species richness higher in the natural wetlands than in the constructed ones, although adult nuisance mosquitoes were associated with all the wetlands. However constructed wetlands may also create or contribute to a mosquito problem. Russell (1999) cautioned that they provided mosquito habitats, and potential for recurrence of diseases that had been eliminated in Australia (e.g., malaria, filariasis). There has been a body of research conducted since then showing that mosquito larvae may be found in the wetlands. For example, Gingrich et al. (2006) found a range of species, including vectors of WNV, in constructed wetlands in Delaware USA. Other researches have focused on management strategies to minimise larval habitats by, for example, vegetation management (Thullen et al. 2002; Jiannino and Walton 2004; Walton and Jiannino 2005). Conversely, flooding felled and dried vegetation, especially *Typha* spp, may increase mosquito abundance (Walton and Jiannino 2005). Marsh design (Walton and Workman 1998; Diemont 2006) or activities such as periodic draining (Mayhew et al. 2004; Dale et al. 2007) indicate that design and maintenance are important for managing mosquito larvae to inhibit adult emergence.

Information gaps

In the area of impacts of mosquito management activities there are knowledge gaps that include: the role of mosquitoes in wetland ecology; long-term impacts on non-target organisms of larvicides; long-term impacts of habitat management on wetlands and, for forested wetlands, the need to identify the nature of larval habitats, suggest management and evaluate impacts.

Restoration

There is a growing literature on restoring damaged wetlands. This is a complex problem and needs an integrated interdisciplinary approach such as in Wassen and Grootjans (1996), who described a key ecohydrology approach in the Netherlands. Keddy (1999), recognising that restoration tended to be somewhat *ad hoc* and based on individual cases, proposed a scheme to unify the process, using assembly rules and indicators of ecosystem integrity, applicable across the range of wetlands from intertidal to fresh. A recent review by Elliott et al. (2007) for estuaries, coastal and marine systems that have been damaged by human land uses, highlighted topics that are poorly understood, such as the ecological goods and services provided by wetlands and they proposed a clarified nomenclature. Zedler and Nelson (2001) wrote a handbook for an adaptive management approach to restoring tidal wetlands (mainly salt marshes), that detailed the practical work involved, for a wide range of physio-chemical and biological variables. Although not necessarily referring to mosquitoes the concepts noted above are useful guides for restoration.

Restoring wetlands that have been affected by mosquito control activities or other disturbance, usually related to modifications such as ditching or impounding, is often in intertidal wetland areas and usually involves restoring tidal circulation. An early approach is exemplified by Broome et al. (1988). Although it does not specifically refer to mosquito-related disturbance, it does discuss water management and that could include mosquito management activities. However, Broome et al. do not consider the interruption to marsh evolution that may have resulted from disturbance over a lengthy period. This is

addressed, for impounded marshes in Florida (USA), by Parkinson et al. (2006, cited above).

Restoration may also specifically include mosquito management. Much of the research on this topic has been conducted in USA. For ditched marshes restoration usually re-introduces tidal flushing by the OMWM approach introduced in the 1960s in New Jersey (USA) and mentioned above (Ferrigno and Jobbins 1968). Tidal flushing has also been restored, or partly restored, to impounded marshes in many areas. A temperate example is in Connecticut (USA), where Swamy et al. (2002) reported a long-term study and noted that recovery may take decades. Recovery may be more rapid in tropical and subtropical areas and research on RIM has been conducted on Florida (USA) marshes that were impounded in the 1950s and 1960s. The method takes into account the need for mosquito management especially during the summer and the need to conserve wetland values. It thus restores tidal access during autumn and winter, and floods the marshes in spring and summer, using floodgates and pumps if needed (Carlson and Carroll Jr. 1983; O'Bryan et al. 1990; Brockmeyer et al. 1997). Lewis and Gilmore (2007) provided a very useful review of mangrove forest restoration in Florida east coast wetlands that had been impounded for mosquito control. They suggested strategies for restoration but did not discuss the potential to also restore the mosquito populations that led to impoundment in the first place.

Restoring wetland function and benefitting wildlife may also have the unintended effect of creating mosquito habitat, though this may be mitigated by vegetation management (removal) (Lawlor et al. 2007). An example from Australia that involved restoring tidal flushing to a wetland showed not only changes to the vegetation with

colonisation by salt marsh and mangrove species (intended) but also increased oviposition by disease vector mosquitoes (unintended) (Turner and Streever 1999).

Information gaps

There is a need for a better understanding of how to restore wetland function without also creating or enhancing mosquito habitat. This requires interdisciplinary approaches, as integrating research from several areas is be relevant to the complex issue of mosquitoes and wetlands. Because ecosystem time frames may be long ones it will require long-term research to assess the effectiveness of restoration as well as monitoring to check for developing mosquito larval habitats.

Impacts of change

While climate change is currently a focus of interest, there are also considerable direct and indirect impacts of human land use decisions on vector-borne disease and mosquitoes and some of these may be indirectly related to climate change. Patz et al. (2004) reviewed the effects of land use changes on a variety of mosquito-borne diseases. They also referred to the millennium Ecosystem Assessment (noted by Pongsiri and Roman (2007), see below).

Specific land use changes such as deforestation can lead to environmental change conducive to mosquito success. Walsh et al. (1993) considered the effects of deforestation on a range of diseases. Previously cited, Yasuoka and Levins (2007) did the same more recently for malaria in Africa and Asia and Vittor et al. (2006) found large increases in anopheline mosquito biting rates in an area that had been deforested in the

Peruvian Amazon. Land use change can have unexpected impacts on mosquito populations. Thus Lindsay et al. (2007), reviewing inland Western Australia, reported an increase in dryland salinity, possibly related to clearing for agriculture, that appeared to have led to the incursion of the salt marsh mosquito (*Aedes camptorhynchus* (Thomson)), a vector of Ross River virus. Derraik and Slaney (2007) reviewed the literature on anthropogenic change (referring also to climate change) and the potential effect on mosquitoes and mosquito-borne disease in New Zealand. Climate change is explored in the next section.

Climate change

In the last quarter of the 20th century climate change was not widely accepted as a fact. With the fourth report of the IPCC (2007) this has changed and climate change and its health impacts are now receiving much attention. The literature that refers specifically to the effect of climate change on mosquito-borne disease goes back to the 1980s when the scenario was uncertain. One of the earliest publications was that of Liehne (1988) who considered the effects in Australia. On the simple expectation, as it then was, that rainfall and temperature would increase, Liehne predicted that malaria risk would increase in the tropical areas and the breeding season of *Culex annulirostris* would extend further south and that this could lead to an increased incidence of polyarthritides diseases such as Ross River virus. He also predicted an increase in the range and frequency of Australian encephalitis disease.

As interest in and concern about climate change and health has increased there has been an increasing number of papers on the topic since the mid 1990s (Fig. 2).

Insert Fig 2

In a relatively early paper Patz and Balbus (1996) foresaw the need to model climate change scenarios and advocated the use of RS and GIS to assess risk related to climate change. Mellor and Leake (2000) reviewed the climatic factors important for vector-borne disease in order to predict climate change effects. Githeko et al. (2000) reviewed the implications of climate change on a regional basis. In summary they expected that the greatest impacts would be felt at the extremes of the temperature ranges at which transmission occurs, via effects on incubation rates and mosquito survival. For specific areas they expected particular outcomes and these are summarised in Table 3 below. They concluded for all areas that the effects of climate change are likely to be variable.

Insert Table 3

Reiter (2001) took a more conservative approach in his review of the history of the major mosquito-borne diseases (malaria, yellow fever and dengue). He noted that “ the natural history of mosquito-borne diseases is complex, and the interplay of climate, ecology, vector biology and many other factors defies simplistic analysis.” (p158). Githeko et al. (2000) and Reiter (2001) both noted that the effects of climate-induced disease risk may be offset by high standards of living. Hunter (2003) reviewed the potential impacts of increased rainfall and higher temperatures on vector-borne disease (and water borne disease). Kuhn et al. (2007) reviewed the use of climate to predict

infectious diseases including mosquito-borne disease, in the light of climate variability and change. That was published prior to the most recent IPCC predictions, but is still relevant to developing early warning systems that would assist disease management and have potential wetland consequences.

Recent papers by Watson et al. (2005), Patz et al. (2005) and Sunyer and Grimalt (2006) discussed the likely impacts of climate change, on vector-borne disease. Watson et al. (2005), in their critical review, noted the need for very broad environmental monitoring and that this should also include responses of biological systems to climate change. A recent paper by McMichael et al. (2006) reviewed the literature on climate change and health at a global level and included a figure that summarised the pathways whereby climate change may affect human health, including the impact on natural systems (and this would include wetlands).

The use of RS, GIS and other climate change modelling approaches is growing. Some predictive modelling has been done on disease risk, based on vector ecology and likely climate changes. An example is the Hotspots models of de Wet et al. (2005) that assessed the risk of Australian disease vector mosquitoes arriving in or dispersing in New Zealand. Zou et al. (2007) used a relatively simple GIS to estimate the risk of WNV; Eisele et al. (2003) focused on malaria in Africa, as did Ceccato et al. (2005). Thomson et al. (1996; 2005) showed that satellite RS, even at low resolution, could assist in predicting malaria outbreaks in Africa, using the Normalised Difference Vegetation Index (NDVI) as an indicator of antecedent rainfall in larval habitats and a precursor to adult mosquito emergence. Other modeling that is relevant includes research on the relationship between mosquitoes and other variables that are likely to be affected by

climate change. An example is research into Ross River virus in Queensland, Australia whereby relationships are shown between the disease and climate (especially rainfall) and tides (sea level) (Kelly-Hope et al. 2004; Tong et al. 2005), climatic variability (Tong and Hu 2002; Tong et al. 2004) or all of these (Naish et al. 2006).

Crane et al. (2005), in the context of chemical control of insects, developed a risk model for climate change and its impacts on insect-borne disease. While it focused on using chemicals to control insects it did so with the objective of protecting aquatic life and hence the approach is relevant to wetland conservation. As with many others, Crane et al. noted the lack of sufficient information about environmental impacts of climate change and the systems that may be affected.

Information gaps

There is information on the likely effects of climate change on mosquitoes and on the diseases they transmit but little that also addresses whole system impacts and management (and by implication wetland issues). As wetlands are likely to be directly and indirectly impacted by climate change there is a need to integrate that information with information on mosquito management.

Balancing the environment and human health issues: adaptive approaches

The issue of human health and biodiversity has been addressed by Pongsiri and Roman (2007), including specific reference to mosquito-borne disease. They reported a Millennium Ecosystem Assessment that showed the need for understanding biodiversity and its relationship with health. Can we manage both for mosquitoes and also sustain the environment? This question was addressed in Dale (1993), with the view that minimal

modification of suitable habitats (salt marshes) is one acceptable solution. Others have also addressed the issue, for example O'Bryan et al. (1990) and Carlson et al. (1999). There are policy documents that outline management from the environmental perspective, for example the recent draft policy for mosquito management in wildlife refuges in USA that acknowledged both the environmental values and the health risks (Department of the Interior 2007).

Adaptive management of wetlands is the subject of research that may be relevant to the interaction between mosquito management and wetland management. McWilliams et al. (2007) used an adaptive approach to impact assessment but also stressed the need to include on-going monitoring, as simple manipulations may have complex results. Olsson et al. (2004) recounted an adaptive management process for managing wetlands in southern Sweden that involved all stakeholders. This is a model that could be used to reconcile what are often seen as conflicting priorities between mosquito control and habitat management. It is similar to the approach of Carlson (1986) describing the Florida Technical Subcommittee on Managed Salt Marshes that was established in 1983 (renamed in 1985) and that resulted in collaboration not confrontation. Other references to adaptive management in a wetland context include La Peyre et al. (2001) and Dale et al. (2006).

There is a risk that wetland management may inadvertently increase mosquito populations. In constructed wetlands the potential is acknowledged and can be avoided by design and maintenance principles. However management in other areas such as in wetland restoration projects may have the unexpected effect of creating conditions suitable for mosquitoes. For example, managing water levels may favour aedine

mosquitoes. An example of water level management without apparent reference to mosquitoes is reported in Palliason et al. (2006) who discuss the effect of water level management to control water lilies in west France and its effect on their species of interest: populations of whiskered tern (*Chlidonias hybridus*). For other examples of conservation focused wetland management research whose results may have mosquito implications, though these are not the subject of the research see Self (2005) and Connor and Gabor (2006).

Information Gaps

In the area of balancing health and wetland interests there is not much that covers both. There has been, and is, a wealth of literature on wetland management that recognises the range of issues that need to be addressed and approaches that can be used (see for example Gosselink et al. 1990; Doust and Doust 1995; Zacharias et al. 2005) including an emphasis on the importance of hydrology, as reviewed in Gilvear and Bradley (2000). Information about mosquitoes and wetlands needs to be integrated into management to help achieve a balance between what may be conflicting interests.

Conclusion

To summarise, wetlands have high intrinsic value and also provide habitats for the immature stages of mosquitoes. Mosquitoes are perceived to be a nuisance and demonstrably transmit diseases. Management of the insect often focuses on the habitats of the immature stages and so wetlands are impacted by mosquito management activities. Information on local mosquito habitats may be held by mosquito managers but little is

published in the refereed scientific literature. Because wetland values are important, mosquito management activities that impact them need to be integrated into overall management plans. There needs to be communication and cooperation between mosquito management and wetland management in order to balance the interests of each and to sustain both wetland and human health, especially in the context of changing land use and climatic environments. As the interaction between wetlands, mosquitoes and mosquito-borne disease is complex there is also a need for interdisciplinary research to provide information that is useful to management.

Finally, a limitation of this review is that it is largely restricted to peer reviewed publications in the English language and thus may have missed important information published in other languages. This is an aspect that could perhaps be explored in the future by an international, multilingual team.

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List of Figure

Fig 1. World distribution of wetlands and some significant mosquito-borne human

diseases (Adapted from the following sources: World wetlands map

<http://soils.usda.gov/use/worldsoils/mapindex/wetlands.html> Accessed 25th October

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Fig 2. Climate change references by year

Tables

Table 1 Global area of wetland types (Source: Schuyt and Brander 2004)

Type	Mangrove	Unvegetated sediment	Salt/ Brackish marsh	Freshwater marsh	Freshwater woodland	Total
Area (1000 ha)	12,112	45,788	6,758	765	9,657	62,967
Total Value \$1000US/yr	185,667	2,848,575	73,382	3,836	333,223	3,444,682

Table 2 Disease, vector and distribution examples by continent/country (as far as possible references include reviews)

Disease	Main vector(s)	Continent/Country and reference
Malaria	Anopheline mosquitoes	General and landuse related (Yasuoka and Levins 2007), Africa, (Githeko et al. 2006); South America, Southeast Asia (Taiwan) (Lien 1991)). Papua New Guinea (Cooper et al. 2006)
Yellow fever	<i>Aedes</i> spp	Africa, S America (Bourgeade and Marchou 2003)
Dengue	<i>Ae. aegypti</i> , <i>Ae. albopictus</i>	General (Bourgeade and Marchou 2003) South America (Mendez et al. 2006), Southeast Asia and Kenya (Hay et al. 2000)
Filariasis	<i>Culex</i> spp	South America (Vezzani et al. 2006), Europe (Cancrini et al. 2006)
Japanese encephalitis	<i>Culex</i> spp	General (Bourgeade and Marchou 2003; Spira 2007), Southeast Asia, Nepal (Partridge et al. 2007)
Ross River or Barmah Forest viruses	<i>Ae. vigilax</i> , <i>Ae. camptorhynchus</i> , <i>Culex annulirostris</i>	Australia (Kelly-Hope et al., 2004), Australia (Quinn et al. 2005; Russell 2002), Fiji (Klapsing et al. 2005)
West Nile virus	<i>Culex</i> spp. <i>Ae. aegypti</i>	North America (Hayes and Gubler 2006; Tyler 2004), South America (Bosch et al. 2007), Europe (Kaptoul et al. 2007), Western Europe (Lundstrom 1999), South Africa (Jupp 2001), India (Paramasivan et al. 2003)

Table 3 Potential impacts of climate change at a regional level (based on Githeko et al. (2006)).

Place	Potential effect
Africa	The response is likely to be variable as the effects of climate change are not evenly spread, nor are the disease implications simple. Research is cited showing that increases in temperature and rainfall have been associated with increased malaria in Kenya, but the same climate effects were accompanied by a decline in malaria in the United Republic of Tanzania.
Europe	Malaria was once common and it could return with climate change and the expanded range of vector habitats. However, the risk is low in the wealthier nations (because of social and health services) though it may be high in impoverished areas.
South America	Malaria and dengue are the major disease risks, though there are cases of yellow fever and encephalitis. The strong effect of El Nino in the equatorial areas is likely to increase the incidence of malaria and dengue and the effects may be exacerbated by any increase in poverty.
North America	Health risks will be affected by the public health and vector control systems that are in place. In southern areas (e.g., Mexico) the infrastructure is less developed and hence there may be increases in diseases such as malaria. They report research that shows that diseases such as St Louis encephalitis may move north as warming occurs.

Asia	Changes to the El Nino effect may lead to increased diseases such as
Australia	malaria in, for example, India. Australia and New Zealand are at risk of
Western	increases in the existing arbovirus diseases such as Ross River and
Pacific	Barmah Forest viruses and Murray Valley encephalitis.
